

## The potential of 2D carbon nitride monolayer as an efficient adsorbent for capturing mercury: A DFT study

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### ABSTRACT

Owing to its large surface area and the close similarity between nitrogen and mercury (Hg) in C<sub>3</sub>N, two-dimensional C<sub>3</sub>N monolayer has the potential of controlling the release of gaseous Hg<sup>0</sup>. Using density functional theory (DFT), we investigated the adsorption and conversion mechanism of Hg species on this monolayer surface. The findings demonstrated that C<sub>4</sub>N<sub>2</sub> site on the surface of the C<sub>3</sub>N monolayer is more capable of adsorbing Hg<sup>0</sup> than C<sub>6</sub> site. Also, the recovery time of the C<sub>3</sub>N monolayer for Hg<sup>0</sup> was approximately 17.63 s and 3.49 μs at 298 and 400 K, respectively. The analyses of electronic structure, charge transport, geometric structure, and adsorption stability demonstrated that the C<sub>3</sub>N monolayer is an encouraging material for Hg<sup>0</sup> adsorption. The C<sub>3</sub>N monolayer could adsorb the oxidizing species of mercury (namely HgBr, HgCl, HgS, and HgO<sub>x</sub>) more easily than Hg<sup>0</sup>, with higher adsorption energy. Overall, the C<sub>3</sub>N monolayer is a promising sorbent material, which is capable of controlling gaseous mercury in an efficient way.

### 1. Introductions

Mercury (Hg) is one of the toxic elements which is released into the water, soil and atmosphere. Also, Hg is accumulated in the food chain, thereby entering the human body. This can present serious threats to ecological environments as well as the human body [1–9]. In 2011, the United States Environmental Protection Agency (EPA) issued the Mercury and Air Toxicity Standard (MATS) with the purpose of reducing the emissions of Hg from power plants, which was implemented in 2012 [10]. China's 12th National People's Congress Standing Committee took a decision in April 2016 on ratifying the Minamata Convention on Hg, urging countries to stop producing, importing and exporting

products which contain Hg for protecting the environment and human health.

Combustion of fossil fuels, especially combustion of coal, is one of the main sources of anthropogenic Hg emission to the atmosphere [11–19]. Wet flue gas desulfurization, fabric filters and electrostatic precipitators can be utilized for removing various Hg forms such as Hg<sup>+1</sup> and Hg<sup>+2</sup> [20–25]. Nonetheless, owing to its insolubility in water, low melting point and high equilibrium vapor pressure, removing Hg<sup>0</sup> is very difficult via air pollution control devices [26–29]. Therefore, one of the main research directions in novel mercury removal technology is to devise an efficient method for removing Hg<sup>0</sup>. In order to remove Hg<sup>0</sup> from flue gas, numerous related removal techniques have been developed. These techniques are based on catalysis [30], ad-

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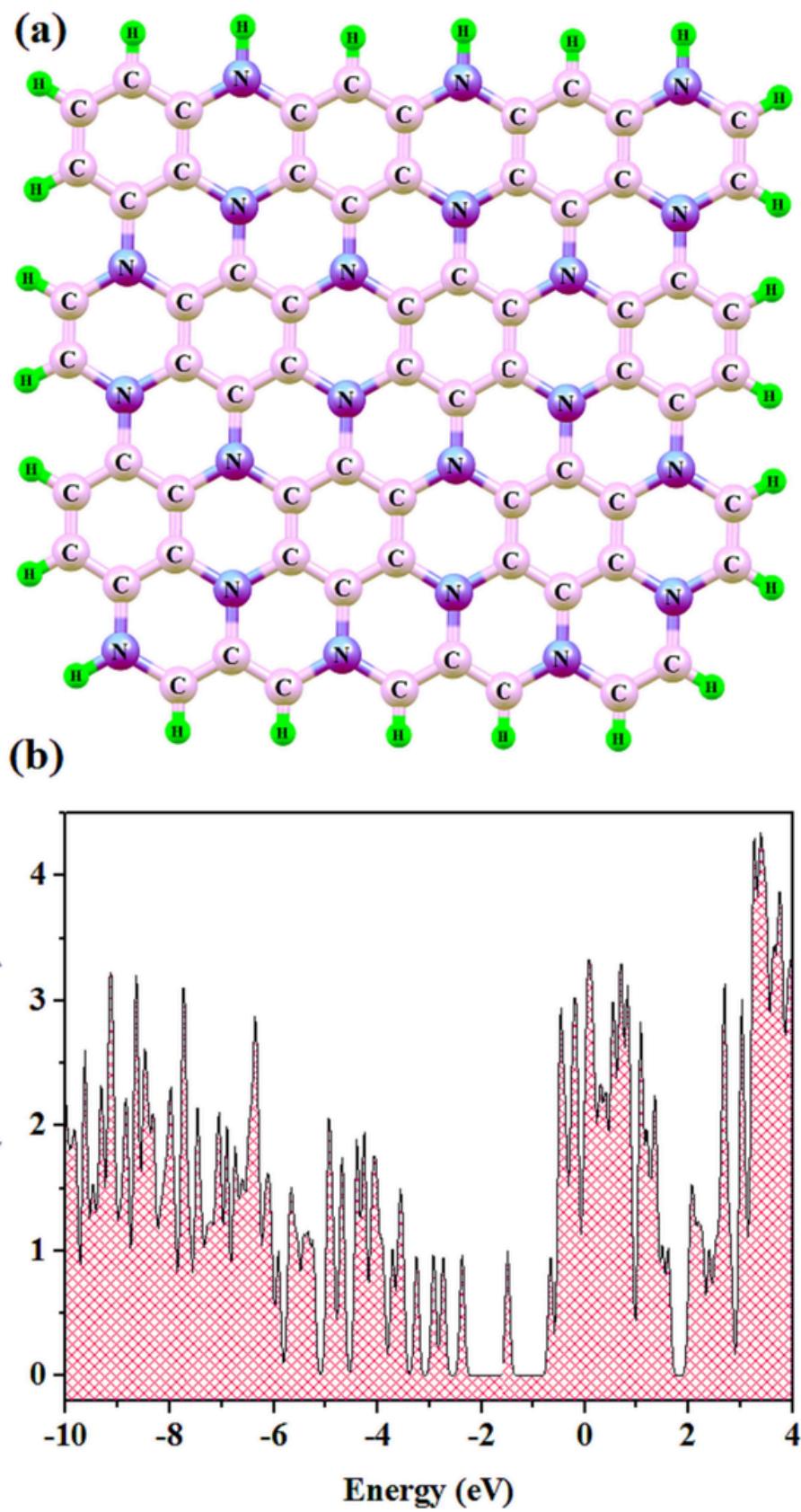


Fig. 1. (a) Optimized structure (top views) and (b) DOS plot of  $\text{C}_3\text{N}$  monolayer.

**Table 1**

The energy and distance of adsorption and Hg<sup>0</sup> Hirshfeld charges at different sites of C<sub>3</sub>N monolayers.

Sites	E <sub>ads</sub> (kJ/mol)	Hirshfeld charge (e)	distances (Å)
C <sub>6</sub> -ring	-19.69	0.109	3.359
C <sub>4</sub> N <sub>2</sub> -ring	-39.84	0.617	2.933

sorption [31–37] and oxidation [38–45]. Since common adsorbents (mineral adsorbents, calcium-based adsorbents, activated carbon, and fly ash) have limitations such as secondary Hg pollution, low efficiency, and high cost, researchers have turned their attention to Hg<sup>0</sup> removal by means of carbon-heteroatom-containing adsorbents [46–51]. We can obtain good adsorption rate and capacity by using carbon-heteroatom-containing adsorbents, which can be efficiently used for Hg removal [52,53].

These days, one of the emerging novel semi-conductors with indirect bandgap (BG) is a two-dimensional (2D) material called C<sub>3</sub>N monolayer, which has carbon and nitrogen atoms [46–49,51,54–57]. By replacing 2 atoms of N onto a 2 × 2 supercell of graphene (Gr), we can produce an intrinsic C<sub>3</sub>N monolayer with 6C atoms and 2 N atoms, which exhibits a planar honeycomb lattice (PHL) with 4.9 Å lattice constant [58]. Replacing N atoms provides the C<sub>3</sub>N monolayer with higher carrier mobility and stronger chemical activity compared to Gr, but structural stability remains the same [59]. The above-mentioned characteristics of the C<sub>3</sub>N monolayer make it possible to use it as a promising sensing material in nanoscale devices. As mentioned above, the PHL of the C<sub>3</sub>N monolayer, which is similar to Gr, provides this material with unique characteristics, including high conductivity, microporous structure, surface area, and unique mechanical characteristics [59], which are essential to a good adsorption material for removing Hg.

Within this study, we scrutinized the adsorption and removal behavior associated with elemental mercury (Hg<sup>0</sup>) on the 2D C<sub>3</sub>N monolayer using DFT calculations and by performing analyses related to reaction energy barrier, charge transport, atomic orbital energy of adsorption, and BG. In order to clarify the removal mechanism of Hg<sup>0</sup> on the C<sub>3</sub>N monolayer, we took the most probable Hg<sup>0</sup> transformation path. In general, the current study provides new insights into removing Hg<sup>0</sup> using the 2D C<sub>3</sub>N monolayer to guide researchers on designing sorbents. The results of this study can guide future research on designing sorbents for the removal of mercury from flue gas, helping to reduce mercury emissions and protect human and environmental health.

## 2. Computational details

The calculations were performed using GAMESS software program with 6-31+G\*\* (d) basis set with the functional Perdew-Burke-Ernzerhof (PBE) [60–65]. For evaluating interactions with long-range and Van der Waals forces have been used DFT-D Grimme's term [66]. Also, the displacement, maximum force, and the energy tolerance accuracy were set to 5 × 10<sup>-3</sup> Å, 2 × 10<sup>-3</sup> Ha/Å and 10<sup>-5</sup> Ha respectively [67–71]. Also, smearing, the global orbital cut-off radius, and self-consistent loop energy were set to 0.005 Ha, 5.0 Å, and 10<sup>-6</sup> Ha, respectively, for computations of static electronic structure for ensuring the correctness of the total energy results [72–75]. Have been performed the Hirshfeld analysis for analyzing the molecular and atomic charge in the systems under study. The charge transport for the adsorption of Hg<sup>0</sup> computing the entire charges of particular molecules which were adsorbed have been defined. A negative and positive charge transport shows the electron acceptor and donor nature of the Hg<sup>0</sup> molecule respectively. For charge transport in Hg<sup>0</sup> adsorption systems, the Mullikan method was used for comparing the results with those of the Hirshfeld method [76].

The energy of adsorption, designated by E<sub>ads</sub>, is one of the significant criteria for investigating the adsorption process mechanism. The

process for the adsorption of molecule X on surface of solid Y is exothermic in nature. E<sub>ads</sub> as follows below:

$$E_{ads} = E_{XY} - (E_X + E_Y) \quad (1)$$

where E<sub>ads</sub> = adsorption energy; E<sub>XY</sub> = total energy after the adsorption of molecule X adsorbed on the surface of solid Y. E<sub>X</sub> and E<sub>Y</sub> = total energy of molecule X and solid adsorbent Y in the ground state. There are two types of adsorption processes, namely physisorption or physical adsorption with E<sub>ads</sub> of less than -50 kJ/mol and chemisorption or chemical adsorption with E<sub>ads</sub> of more than -50 kJ/mol.

The difference in electron density ( $\Delta\rho$ ) have been calculated as follows:

$$\Delta\rho = \rho_{XY} - \rho_X - \rho_Y \quad (2)$$

where  $\rho_{XY}$  = total electron density of the adsorption system.  $\rho_X$  = unperturbed electron densities of surface X; and  $\rho_Y$  = isolated adsorbate molecule Y.

## 3. Results and discussion

### 3.1. Structure and the electronic parameters of the C<sub>3</sub>N monolayer

According to Fig. 1(a), the C<sub>3</sub>N monolayer includes a PHL with six C atoms and 2 N atoms. In this monolayer, σ and π bonds are formed by the sp<sup>2</sup> hybridization of C and N atomic orbitals. The length of C—C and C—N bonds are 1.42 Å and 1.401 Å respectively, in line with values reported in the literature [77]. As illustrated in Fig. 1 b, we selected an orthogonal C<sub>3</sub>N monolayer with a BG of 0.89 eV, in line with the value reported in the literature [77]. The density of states (DOS) plot of pristine C<sub>3</sub>N monolayer demonstrates the peaks of fermi level at -1.96 eV. We conducted the Bader charge analysis for computing the charge transport [78]. In fact, each N atom received 1.26 e electrons from its adjacent C atoms because the electronegativity of N (3.04) is more than that of C (2.55) (Table 1).

### 3.2. Hg<sup>0</sup> adsorption on the C<sub>3</sub>N monolayer

One of the important steps in the conversion of heterogeneous Hg is the adsorption of Hg<sup>0</sup> on the surface of the C<sub>3</sub>N monolayer. Therefore, first investigated the adsorption behavior of Hg<sup>0</sup> on the C<sub>3</sub>N monolayer. As illustrated in Fig. 2, there were two possible sites for the adsorption of Hg<sup>0</sup> on the surface of the C<sub>3</sub>N monolayer. Hg<sup>0</sup> was adsorbed at the N atom of the C<sub>4</sub>N<sub>2</sub> ring site with higher E<sub>ads</sub> of -39.84 kJ/mol than that of the C atom of C<sub>6</sub> ring site with E<sub>ads</sub> of -19.69 kJ/mol. Additionally, at C<sub>4</sub>N<sub>2</sub> site, the distance between Hg atom and N atom was 2.933 Å, which was shorter than that between Hg atom and the C atom at C<sub>6</sub> site that was 3.359 Å. Based on the results, Hg atom had a physical adsorption at C<sub>6</sub> site, but it had a strong physical adsorption at C<sub>4</sub>N<sub>2</sub> site on the surface of the C<sub>3</sub>N monolayer. In the substituted C<sub>4</sub>N<sub>2</sub> ring, a bond between Hg and the remaining N was formed because the other N is absent, which was capable of facilitating the adsorption of Hg atom. As illustrated in Fig. 3, have been performed the analysis of charge density difference to further assess the charge transport in the stable adsorption system. Based on the results, there was a charge redistribution in the adsorption of Hg<sup>0</sup> on the C<sub>3</sub>N monolayer surface. In general, there was an accumulation of electrons near the C atom and there was a depletion of electrons near Hg and N atoms. The considerable transport of electrons confirmed that the Hg atom and N atom interacted at C<sub>4</sub>N<sub>2</sub> site. Based on the Hirshfeld analysis [79], the amount of electron transport for Hg was more at C<sub>4</sub>N<sub>2</sub> site (0.617 e) than the amount of electron transport at C<sub>6</sub> site (0.109 e). Overall, the findings demonstrated that the efficiency of C<sub>4</sub>N<sub>2</sub> site is more than that of C<sub>6</sub> site in adsorbing Hg. The value of adsorption energy reported in the literature for other com-

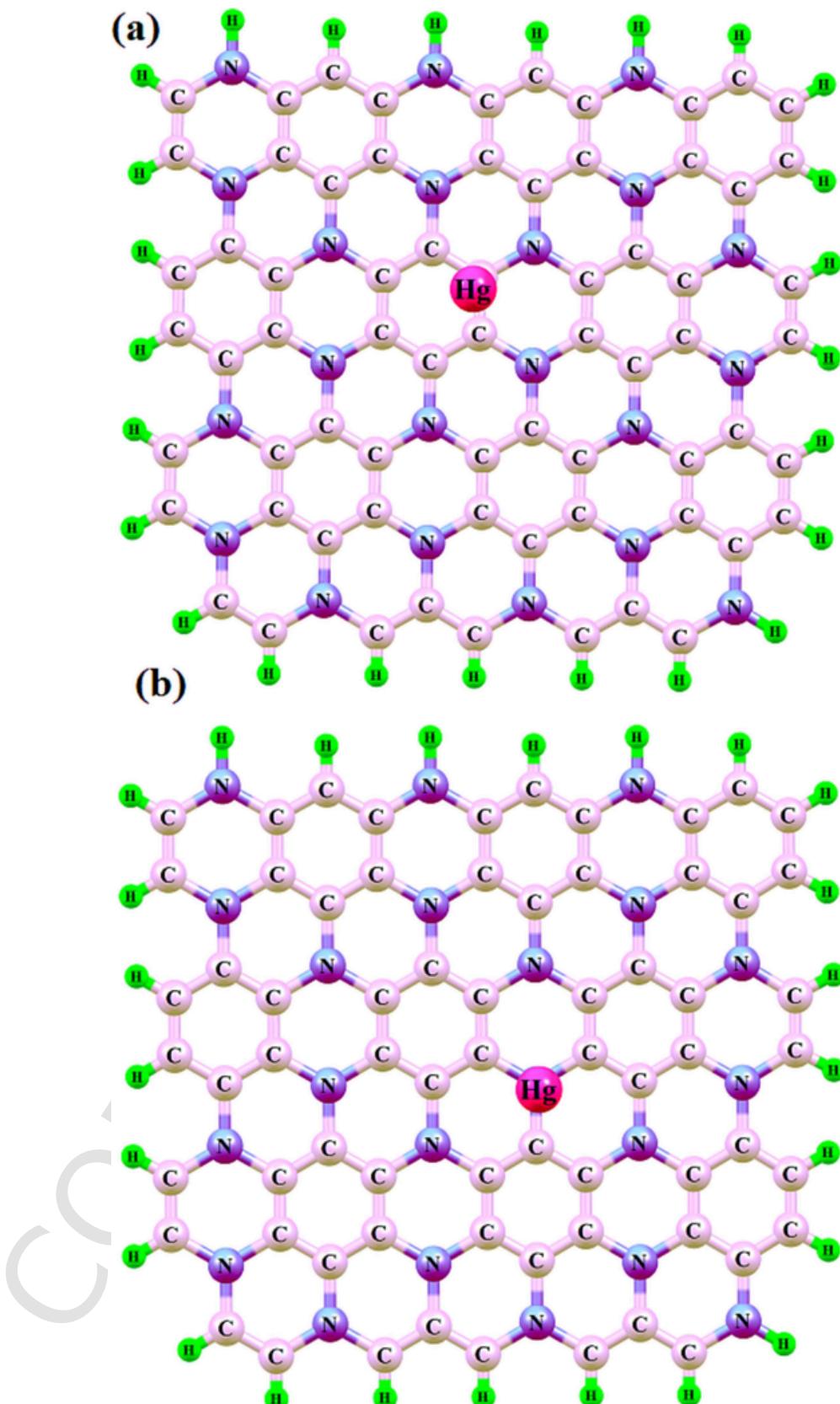
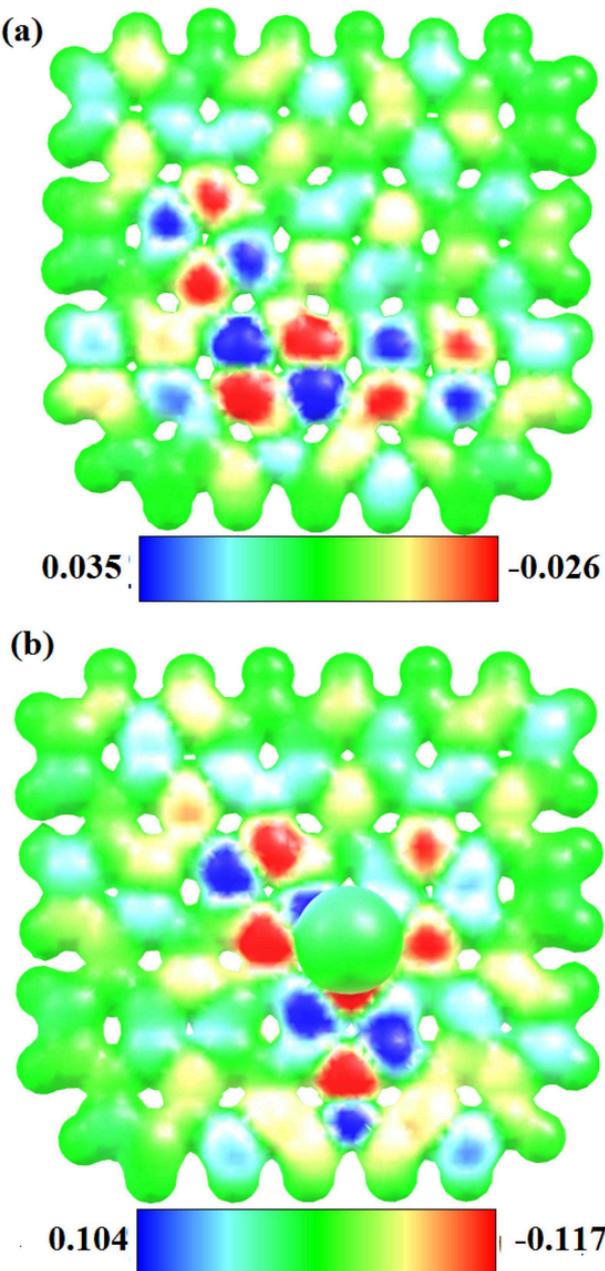


Fig. 2. Optimized structures for adsorption of  $\text{Hg}^0$  at sites (a)  $\text{C}_6$  and (b)  $\text{C}_4\text{N}_2$  on the  $\text{C}_3\text{N}$  surface.



**Fig. 3.** Map of the charge density difference for the  $C_3N$  monolayer (a) before and (b) after adsorption of  $Hg^0$ .

pounds on the  $C_3N$  surface shows that the Hg absorption energy is higher than carbon dioxide [80], nitrogen dioxide, and sulfur dioxide [81].

### 3.3. Electronic characteristics

Using the DOS diagrams in Fig. 4, which was plotted after the adsorption, the interaction mechanism between the  $Hg^0$  atoms at  $C_4N_2$  site on the surface of the  $C_3N$  monolayer have been understand. Fig. 4 depicts the DOS diagrams of the  $C_3N$  monolayer after the adsorption of the  $Hg^0$  molecule. Due to the charge transport from  $Hg^0$  to the surface of the  $C_3N$  monolayer, all the orbitals of  $Hg^0$  atom moved to levels with low energy after  $Hg^0$  was adsorbed on the surface of the  $C_3N$  monolayer. At  $C_4N_2$  sites, the hybridization of N atom can be observed in the s- / p orbitals between  $Hg^0$  and the neighboring N atom close to the Fermi level ( $-1.78$  eV). In agreement with previous results, a strong or-

bital hybridization resulted in a strong interaction for the  $Hg^0$  adsorption on the monolayer surface.

Also was investigated the BG to expound on the difference between the two adsorption sites. The BG of the pristine  $C_3N$  monolayer was  $0.89$  eV, demonstrating that the energy needed to transport electrons to the conduction band (CB) from the valence band (VB) was low. Also, following the adsorption of  $Hg^0$  on the monolayer surface, BG decreased from  $0.89$  eV to  $0.65$  eV. The adsorption of  $Hg^0$  at  $C_4N_2$  site resulted in a significant decrease in the BG, facilitating the transport of electrons from the VB to the CB. These results can be used as a signal for the adsorption of  $Hg^0$ .

### 3.4. Evaluation of $C_3N$ monolayer recovery time (RT)

As discussed above, the  $C_3N$  monolayer is a good adsorbent for  $Hg^0$  due to its significant adsorption energy. However, the reusable characteristic of the  $C_3N$  monolayer should also be taken into account in order to consider the  $C_3N$  monolayer as a good adsorbent, which can be assessed by the RT. The RT in keeping with the transition state theory as follows [82]:

$$\tau = v_0^{-1} \exp\left(\frac{-E_{ads}}{k_B T}\right) \quad (3)$$

where  $v_0$  = attempt frequency ( $10^{-13}$  s $^{-1}$ ); and  $E_{ads}$  = adsorption energy [83]. Here,  $k_B$  = Boltzmann constant; and  $T$  = temperature (K). At ambient temperature (298 K), the RT for the desorption of  $Hg^0$  from the surface of the  $C_3N$  adsorbent was  $17.63$   $\mu$ s. Moreover, at higher temperature than 298 K, i.e., 400 K, the RT was computed to be  $3.49$   $\mu$ s on the  $C_3N$  monolayer, demonstrating the reusable characteristic of the  $C_3N$  monolayer under humid conditions and/or at higher temperatures than 298 K. Therefore, due to its excellent stability and short RT, have been considered the  $C_3N$  monolayer a promising  $Hg^0$  adsorbent.

### 3.5. Adsorption of various Hg species on the $C_3N$ monolayer

The Hg removal mechanism is dependent on the Hg species which exists in industrial flue gas. Hence, explaining the adsorption behavior of different species of Hg is of paramount importance in gaining a deeper understanding of its conversion mechanism. There are two important types of oxidized species of Hg, namely mercury sulfide ( $HgS$ ) and mercury oxide ( $HgO$ ). However, two important intermediate materials in the Hg conversion process are  $HgBr$  and  $HgCl$  [34]. Therefore, for further investigated how the adsorption ability of the  $C_3N$  monolayer (defective and clean) was affected by different species of Hg such as  $HgBr$ ,  $HgCl$ ,  $HgS$ ,  $HgO$ , and  $Hg^0$ . Fig. 5 and Fig. 1S shows the obtained  $E_{ads}$  on the surface of the  $C_3N$  monolayer. Each Hg species has a different  $E_{ads}$  on the  $C_3N$  monolayer surface, which shows that there is a difference in the bonding strength between different Hg species and the  $C_3N$  monolayer. Overall,  $E_{ads}$  of the different species of Hg is in the following order:  $HgCl > HgBr > HgS > HgO > Hg^0$ . This show that the  $C_3N$  monolayer can capture Hg in its oxidation form more easily. Strong physical adsorption is the main cause for the adsorption of different species of Hg on the  $C_3N$  monolayer rather than physisorption (or Van der Waals force). Furthermore, the charge transport in the Hg atom of  $HgO$  ( $0.873$  e),  $HgCl$  ( $0.664$  e) and  $HgBr$  ( $0.703$  e) was more, demonstrating the strong physical adsorption of oxidized species of Hg on the surface of the  $C_3N$  monolayer.

## 4. Conclusion

Using DFT, the adsorption and conversion mechanism of  $Hg^0$  on the 2D  $C_3N$  monolayer was investigated and expounded on. Based on the findings,  $Hg^0$  had a physical adsorption at  $C_6$  site, while it had a strong physical adsorption at  $C_4N_2$  site on the surface of this monolayer. Obvi-

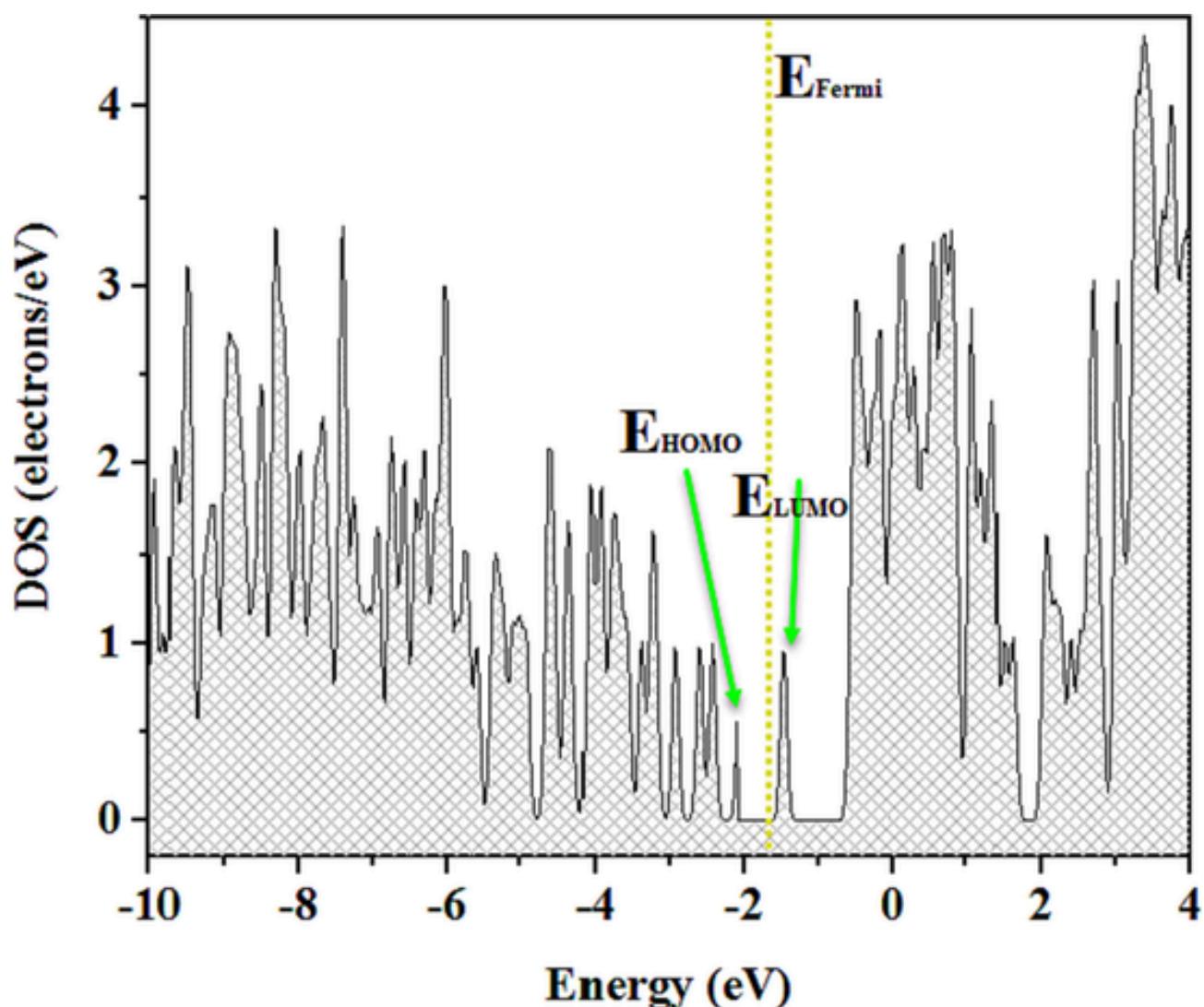


Fig. 4. DOS plot of  $\text{C}_3\text{N}$  monolayer after adsorption  $\text{Hg}^0$  (the dash line is Fermi level).

ously, the considerable charge transport and orbital hybridization resulted in a strong interaction between  $\text{Hg}^0$  and the  $\text{C}_3\text{N}$  monolayer. The RT of the  $\text{C}_3\text{N}$  monolayer for  $\text{Hg}^0$  was approximately 17.63 s and 3.49  $\mu\text{s}$  at different temperatures of 298 and 400 K respectively. Therefore, the  $\text{C}_3\text{N}$  monolayer was considered a promising adsorbent material with noticeable recovery performance for  $\text{Hg}^0$  adsorption. Also, compared to  $\text{Hg}^0$ , oxidized Hg species tend to be adsorbed more by the  $\text{C}_3\text{N}$  monolayer with higher  $E_{\text{ads}}$ . The current work suggests that the  $\text{C}_3\text{N}$  monolayer is a good adsorbent, capable of controlling gaseous Hg in flue gas in an efficient way.

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#### Ethical approval

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#### CRediT authorship contribution statement

**M. J. Saadh, D. Singh, D. Mayorga, A. Kumar, M. Albuja, A. I. Saber:** Methodology, Software, Writing - review & editing; Management and responsibility for the research activity planning and execution; **M. A. Mustafa:** Conceptualization, Methodology, Software, Writing, Conceptualization, Methodology; **Y. Elmasry, N. Sun:** Writing - original draft, Methodology, Software, review & editing.

#### Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

#### Data availability

No data was used for the research described in the article.

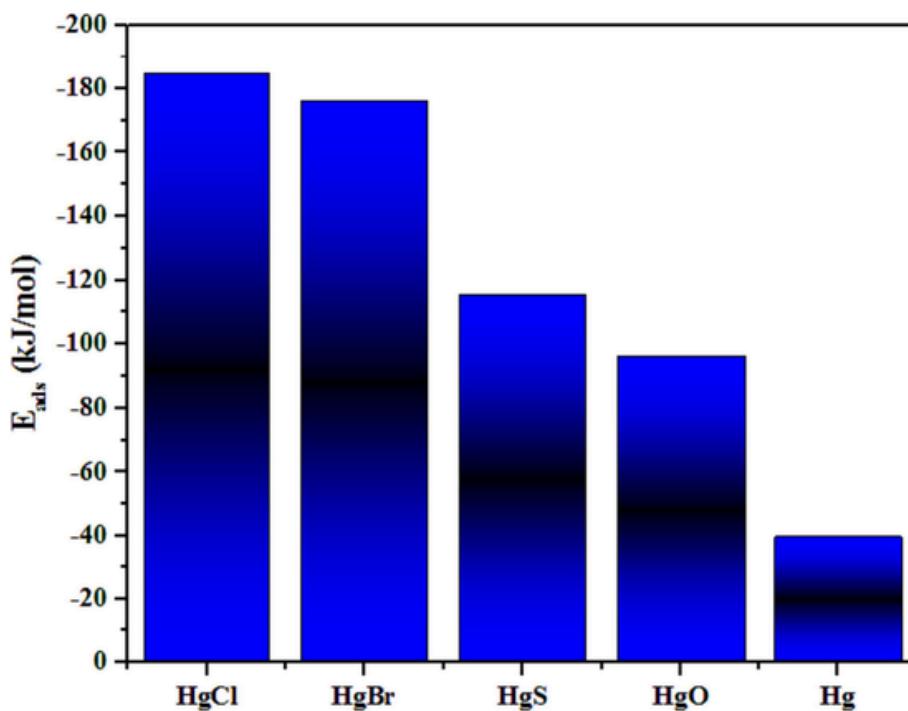


Fig. 5. Comparison of the energy of adsorption for different Hg species on the  $\text{C}_3\text{N}$  monolayer.

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