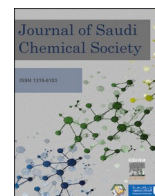


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Novel magnetically separable g-C₃N₄/TiO₂/CuFe₂O₄ photocatalyst for efficient degradation of tetracycline under visible light irradiation: Optimization of process by RSM

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ABSTRACT

Herein, a novel magnetic visible-driven g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite with excellent photocatalytic performance was successfully prepared and employed for photodegradation of tetracycline. Several analysis including X-Ray diffraction (XRD), Fourier Transform Infrared Spectroscopy (FTIR), Field Emission Scanning Electron Microscopy (FESEM), energy dispersive X-ray (EDX), Vibrating-Sample Magnetometer (VSM), and Ultraviolet-Visible Diffuse Reflectance Spectroscopy (UV-Vis DRS) were performed in order to study the structural, optical, magnetic, as well as morphological properties of nanocomposite. The optical band gap of g-C₃N₄/TiO₂/CuFe₂O₄ heterostructure was found to be red shifted to 2.45 eV from 3.15 eV for pure TiO₂. Enhanced separation of photoinduced electron-hole pairs and enhanced visible light absorption capacity of nanocomposite lead to a maximum tetracycline photodegradation efficiency. Response surface methodology (RSM) was used to investigate the influence four independent variables, including initial photocatalyst dosage (7–14 g/L), TC concentration (20–30 ppm), solution pH (5.5–7.5), and irradiation time (20–40 min), and optimize the TC degradation efficiency. The g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite was able to separate and recycle easily using an external magnetic field, and the results of reusability was shown its high stability after 5 cycles. Active species trapping experiments suggested that holes and hydroxyl radicals played a crucial role in the TC degradation process. Finally, a potential photocatalytic mechanism for photodegradation of TC was proposed.

1. Introduction

The pollution of water with pharmaceutical antibiotics such as tetracycline (TC) has attracted much attention due to their poor metabolism, high chemical stability, and high water-solubility [1]. These

pollutants even in the low concentration can lead to serious problems such as drug resistance and chronic toxicity, which is a considerable threat for human health and aquatic environment [2,3]. Therefore, the eliminating of such pollutants from wastewater is highly desirable. A variety of water purification techniques such as biodegradation [4],

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chlorination [5], flocculation [6], and adsorption [7,8,9] have been used for eliminating these toxic pollutants. Although somewhat efficient, the application of these techniques is limited due to numerous shortcomings such as generation byproducts, high cost, time-consuming, and incomplete elimination. To solve these drawbacks, advanced oxidation processes (AOPs), which use reactive oxygen species to remove the pollutants have been attracted more attention [10,11]. Photocatalyst based semiconductors are one of the most efficient AOPs techniques for wastewater treatment, due to their high efficiency and utilize solar energy directly [12–15].

TiO₂ have been regarded as the promising photocatalyst material for removal of pollutants from aqueous solution because of its high stability, nontoxicity, excellent physicochemical properties, and low-cost [16,17]. However, its performance is hindered by the wide band gap energy and rapid recombination of photoinduced carriers [18,19]. Hence, numerous methodologies have been developed to improve sun light photocatalytic performance of TiO₂. One of the promising strategy is coupling of TiO₂ with other semiconductors and the construction of heterojunction nanocomposites [20–22]. Among these semiconductor materials, graphitic carbon nitride (g-C₃N₄) is commonly used owing to its nontoxicity, narrow band gap, cost-effectiveness, and high stability [23,24]. These interesting properties make g-C₃N₄ very useful in various field of science such as solar cell [25], chemical sensors [26], and photoelectronic [27], and photocatalyst [28–30]. Nevertheless, the photocatalytic functionality of pure g-C₃N₄ is hindered by the rapid recombination kinetics of photoexcited charge carriers, alongside its restricted surface area. These constraints impede the material's overall photocatalytic efficiency, necessitating further enhancement strategies to optimize its performance in catalytic applications. Gündoğmuş et al. prepared TiO₂/g-C₃N₄ heterojunction through one pot hydrothermal procedure, which exhibited superb photodegradation performance to methylene blue dye under irradiation of visible light [31]. Preparation and application of magnetic photocatalysts with interesting properties such as fast separation, low cost, and basic technical necessities are promising procedure in the wastewater treatment [32]. Thus, the photocatalysts can be coupled with magnetic nanoparticles to facilitate separation.

Recently, the spinel ferrite structure (MFe₂O₄ like CuFe₂O₄, MnFe₂O₄, and CoFe₂O₄, etc.) photocatalysts were widely explored to degradation of organic pollutants. Among them, copper ferrite (CuFe₂O₄) has been demonstrated as a promising ferromagnetic materials, as it possess many advantages such as simple preparation, high chemical and thermal stability, low toxicity, narrow band gap, low cost, and fascinating magnetically separable properties [33,34]. Therefore, the coupling CuFe₂O₄ with other materials provides an efficient and convenient strategy for increased visible-light photocatalytic performance, which can be easily separated using magnetic field. Atacan and co-workers were synthesized CuFe₂O₄/MoS₂ p-n heterojunction and used for photodegradation of Rhodamine B [35]. The synthesized magnetic nanocomposite was shown excellent photocatalytic performance and reusability.

In this research, we reported preparation of a novel magnetic recoverable g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite by a simple wet impregnation procedure. The as-prepared materials were characterized by different analyses such as XRD, FT-IR, FESEM, EDX, VSM, and UV–vis DRS. The photocatalytic activity of the as-fabricated nanocomposite was evaluated using studying the photodegradation of TC and shown photoinduced increased photocatalytic activity. Besides, the effect of TC concentration, photocatalyst dosage, time of irradiation, and solution pH on the photodegradation process were investigated using response surface methodology (RSM). The as-prepared nanocomposite exhibited good reusability against TC for continuous cyclic runs. Scavenging experiments was then used to identify primarily reactive species. Furthermore, the possible degradation pathway of TC by g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite was investigated in detail.

2. Experimental

2.1. Chemicals and reagents

Titanium (IV) isopropoxide (TTIP), melamine (C₃H₆N₆), copper nitrate trihydrate (Cu(NO₃)₂·3H₂O), Iron (III) nitrate nonahydrate (Fe(NO₃)₃·9H₂O), isopropanol (C₃H₈O), sodium hydroxide (NaOH), nitric acid (HNO₃), and methanol (CH₃OH) were the reagents which used for synthesis TiO₂, g-C₃N₄, CuFe₂O₄, g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite. All the reagents were applied without further purification.

2.2. Synthesis of TiO₂

The TiO₂ nanoparticles were prepared through hydrolyze and peptization of TTIP solution according to the reported procedures. Briefly, 5 mL TTIP and 15 mL isopropanol were slowly added to 250 mL of distilled water (pH = 3, adjusted by adding HNO₃). The mixture was magnetically stirred at room temperature for 3 h, and after that kept at 70 °C for 22 h in an oven. Finally, the resulted product was separated by centrifugation and dried at 70 °C, after which the sample was calcinated or annealed at 400 °C for 2 h.

2.3. Synthesis of g-C₃N₄

For preparation of g-C₃N₄ nanosheets at first a certain amount of melamine was put in a crucible and heated at 550 °C for 4 h in a muffle furnace. Then, the resultant yellow sample grounded and ultrasonicated in methanol for 30 min. Finally, the product was washed with distilled water and dried at 65 °C for 12 h.

2.4. Synthesis of CuFe₂O₄

Magnetic CuFe₂O₄ nanoparticles was prepared by co-precipitation procedure [36]. Typically, 5 mmol Fe(NO₃)₃·9H₂O, and 2.5 mmol Cu(NO₃)₂·3H₂O were dissolved in distilled water (100 mL) magnetically stirring. After that, 75 mL NaOH (4 M) was slowly dropped into the obtained mixture and kept at 90 °C for 2 h. The obtained magnetic CuFe₂O₄ nanoparticles were collected by an external magnet, washed with distilled water, and finally dried at 75 °C.

2.5. Synthesis of TiO₂/g-C₃N₄/CuFe₂O₄

TiO₂/g-C₃N₄/CuFe₂O₄ was prepared using a simple wet impregnation procedure. The equal weight ratio (1:1:1) of TiO₂, g-C₃N₄, and CuFe₂O₄ were dispersed in 30 mL EtOH individually, ultrasonicated for 30 min, and then mixed together. The obtained mixture was magnetically stirred for 24 h. Finally, the obtained magnetic nanocomposite were isolated by external magnet, and dried at 80 °C.

2.6. Characterization techniques

The crystal structure and phase purity of the as-prepared samples were analyzed by XRD technique on an MACSCIENCE MXP3 diffractometer under Cu K α radiation. The surface functional group of the as-synthesized samples were obtained using a Nicolet Avatar 370 spectrometer. Using a ZEISS SIGMA VP scanning electron microscope, the samples went through FESEM and EDS analyses to study their surface morphology, elemental constitution, and distribution. JASCO V-65 spectrophotometer was used to record the UV–visible absorption and diffuse reflectance (DRS) spectra of the as-prepared samples. The magnetic properties of the as-prepared g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite were analyzed with Lakeshore VSM 7410 magnetometer.

2.7. Assessment of photocatalytic activity

The photocatalytic performance of the as-synthesized samples were

Table 1
Range and level values of independent variables.

Parameter	Code	Real values				
		$-\alpha$	-1	0	$+1$	$+\alpha$
Dosage (mg)	A	3.5	7.0	10.5	14.0	17.5
Initial concentration (ppm)	B	15	20	25	30	35
Contact time (min)	C	10	20	30	40	50
pH	D	4.5	5.5	6.5	7.5	8.5

Table 2
Design of experiments matrix along with experimental values of removal percentage of TC by the $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$.

Run	A (mg)	B (ppm)	C (min)	D:pH	%R
1	10.5	25	30	8.5	73.26
2	14	30	20	5.5	84.26
3	10.5	25	30	6.5	67.21
4	7	30	40	7.5	70.12
5	10.5	15	30	6.5	91.28
6	10.5	25	30	6.5	60.18
7	10.5	25	30	6.5	62.29
8	7	20	20	7.5	69.51
9	7	20	40	7.5	77.81
10	7	20	40	5.5	86.51
11	7	30	20	5.5	70.12
12	3.5	25	30	6.5	59.68
13	10.5	35	30	6.5	72.25
14	10.5	25	30	4.5	91.02
15	10.5	25	30	6.5	63.18
16	14	30	20	7.5	76.02
17	10.5	25	10	6.5	68.12
18	14	20	40	5.5	98.42
19	7	30	20	7.5	65.13
20	14	20	40	7.5	90.08
21	14	20	20	7.5	83.25
22	14	30	40	5.5	90.09
23	14	30	40	7.5	82.18
24	10.5	25	30	6.5	61.89
25	7	30	40	5.5	77.99
26	17.5	25	30	6.5	89.11
27	10.5	25	50	6.5	85.18
28	7	20	20	5.5	77.81
29	14	20	20	5.5	92.31
30	10.5	25	30	6.5	64.05

investigated using photodegradation of TC under irradiation of visible light using LED lamp. In a typical procedure, a certain quantities of photocatalyst was dispersed into TC solution with desired

concentration, and then stirred in the dark for 30 min to get adsorption/desorption equilibrium. Then, the suspension was exposed to visible light for a specific time. The residual concentration of TC was quantified through the utilization of UV-visible spectrophotometry ($\lambda_{\text{max}} = 365$ nm).

2.8. Experimental design

In order to evaluate the effects of key experimental variables on the degradation efficiency and also optimize the photocatalytic process, RSM based central composite design (CCD) was performed. Design Expert 10 software was used to design of experimental. In this research, four key variables containing dosage of $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ (mg, A), initial concentration of TC ($\text{mg}\cdot\text{L}^{-1}$, B), irradiation time (min, C), and solution pH (D) were considered as significant experimental variables affecting the photocatalytic performance. Table 1 was listed the ranges and levels of the four selected operational variables. Based on CCD-RSM a set of 30 experiment were conducted and the obtained matrix was shown in Table 2.

3. Results and discussion

3.1. Characterization of photocatalyst

The XRD crystalline patterns of TiO_2 , $g\text{-C}_3\text{N}_4$, CuFe_2O_4 , and $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ nanocomposite were depicted in Fig. 1a. The pattern of TiO_2 was shown diffraction peaks at $2\theta = 25.55^\circ$, 37.94° , 48.28° , 55.0° , 63.10° , 70.02° , and 75.6° with corresponding hkl planes of (101), (004), (200), (211), (204), (220), and (215), respectively, which was well matched with anatase phase of TiO_2 (JCPDS No. 00-021-1272). For pure $g\text{-C}_3\text{N}_4$, the diffraction peaks at $2\theta = 13.07^\circ$ and 27.72° were correspond to the (100), and (002) plans (JCPDS No. 87-1526). Furthermore, the XRD pattern of CuFe_2O_4 was displayed peaks at 2θ of 18.5° , 30.2° , 35.8° , 43.03° , 57.05° and 62.8° , which are related to (111), (220), (311), (400), (511) and (440) planes, respectively. These reflections are agree well with spinel cubic phase of CuFe_2O_4 (JCPDS No. 25-0283). The characteristic peaks of TiO_2 , $g\text{-C}_3\text{N}_4$, and CuFe_2O_4 coexisted in the pattern of $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ nanocomposite, which confirm the TiO_2 , and CuFe_2O_4 were successfully anchored on the surface of $g\text{-C}_3\text{N}_4$ nanosheets. Furthermore, in the XRD pattern of $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ nanocomposite no additional diffraction peaks were observed, which confirm the phase purity of the nanocomposite. Thus, TiO_2 , $g\text{-C}_3\text{N}_4$, and CuFe_2O_4 coexist stably in the ternary heterojunction without chemical reaction. In addition the XRD patterns

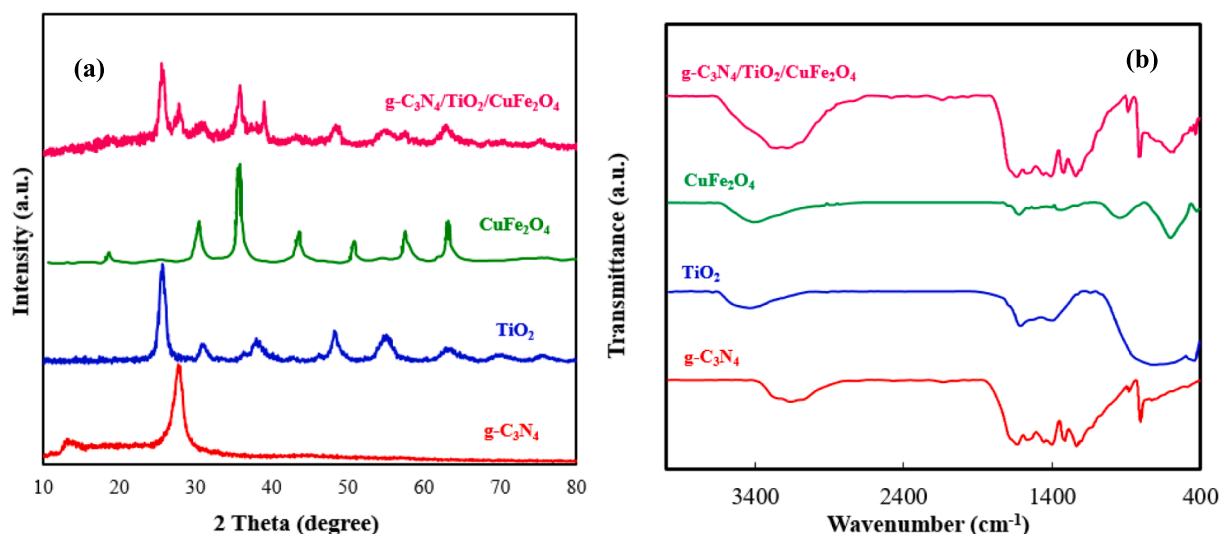


Fig. 1. XRD patterns (a), and FTIR spectra (b) of $g\text{-C}_3\text{N}_4$, TiO_2 , CuFe_2O_4 , and $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ nanocomposite.

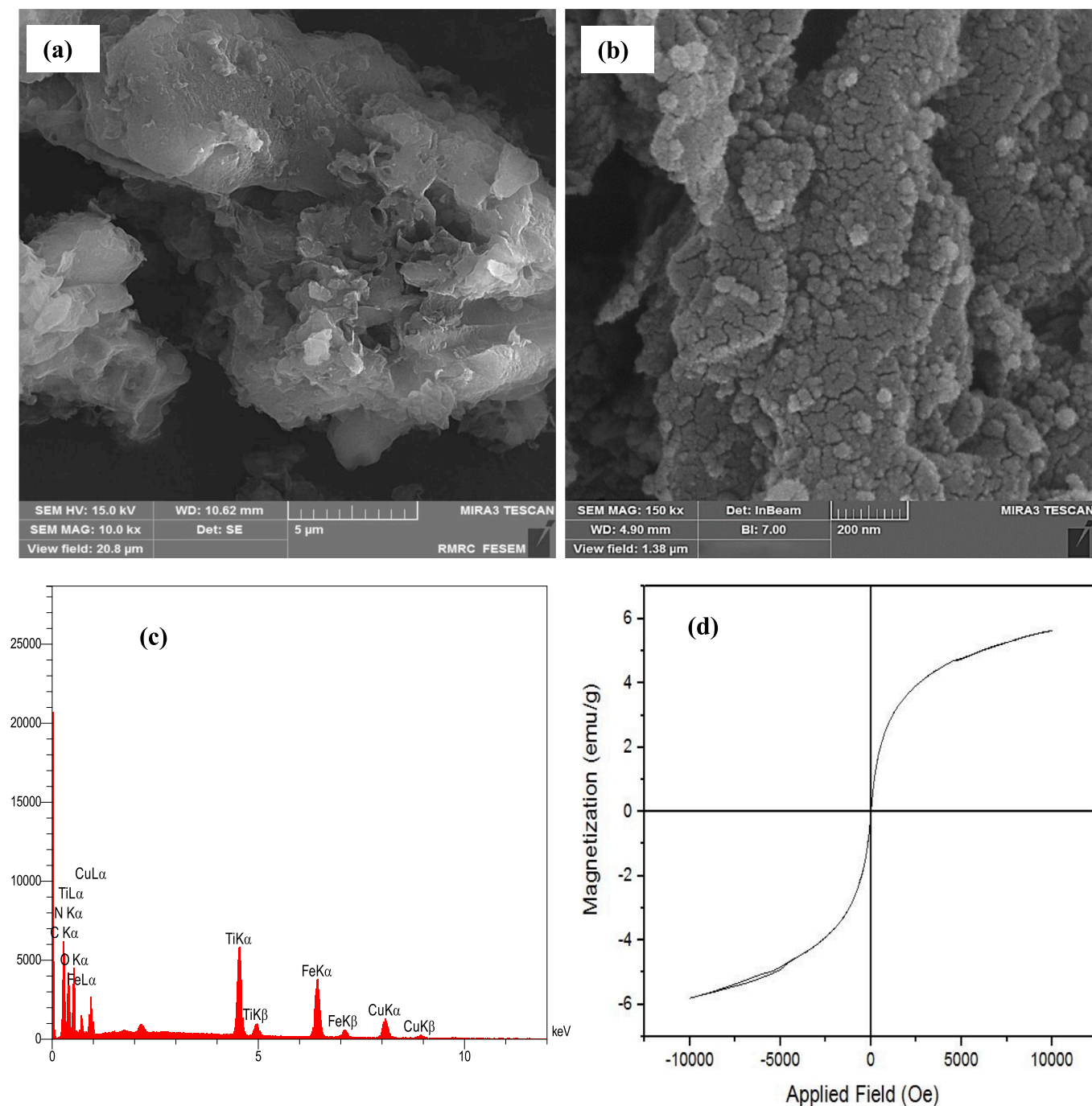


Fig. 2. FE-SEM image g-C₃N₄ (a), and g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite (b), EDS spectrum, of g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite (c), and magnetization curve of g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite (d).

of g-C₃N₄/TiO₂, TiO₂/CuFe₂O₄, and g-C₃N₄/CuFe₂O₄ confirm the successful synthesis of these nanocomposites (Fig. S1).

The FTIR technique was used to determine the surface functional group of the samples and the obtained spectra were displayed in the Fig. 1b. In the spectra of all the synthesized samples, the absorption bands at around 3320 and 1650 cm⁻¹ are corresponded to stretching and bending vibrations of OH groups, respectively. In the case of TiO₂, the strong absorption bands at 500–800 cm⁻¹ are related to Ti–O and Ti–O–Ti bonds stretching vibration. For pure CuFe₂O₄, the absorption peaks at 450 and 580 cm⁻¹ are related to the stretching vibrations of metal–oxygen within the octahedral (Cu–O) and tetrahedral (Fe–O) sites in spinel structures, respectively. The FTIR spectrum of g-C₃N₄ has

revealed absorption bands at 1239 cm⁻¹, 13121 cm⁻¹, 1411 cm⁻¹, and 1578 cm⁻¹ which were related to the C–N aromatic stretching vibration. Furthermore, the peaks around 807, and 1637 cm⁻¹ are related to the tri-s-triazine units and C=N stretching vibrations. Moreover, all the characteristic absorption bands of TiO₂, g-C₃N₄, and CuFe₂O₄ were coexisted in the FTIR spectrum of g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite, which confirmed the successful preparation of the ternary nanocomposite.

The morphology and microstructure of the synthesized g-C₃N₄ as well as g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite were analyzed by FESEM and the images were depicted in Fig. 2. According to the Fig. 2a, g-C₃N₄ was shown the layered nanostructures containing nanosheets with wrinkle. Fig. 2b was exhibited that the TiO₂ nanoparticles and CuFe₂O₄

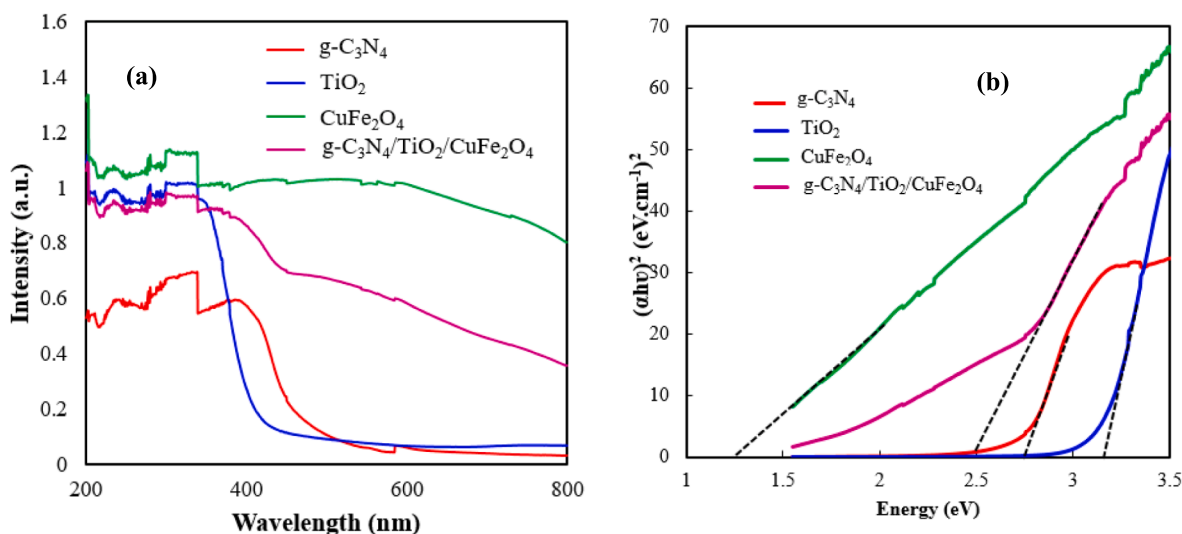


Fig. 3. Diffuse reflectance spectra (a), Tauc's plots (b) of g-C₃N₄, TiO₂, CuFe₂O₄, and g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite.

Table 3
ANOVA results of the quadratic model.

Source	Sum of Squares	df	Mean Square	F Value	p-value	
Model	3679.47	14	262.82	57.48	< 0.0001	significant
A-Dosage	1072.94	1	1072.94	234.66	< 0.0001	
B-Concentration	398.94	1	398.94	87.25	< 0.0001	
C-Time	329.37	1	329.37	72.04	< 0.0001	
D-pH	407.80	1	407.80	89.19	< 0.0001	
AB	0.65	1	0.65	0.14	0.7110	
AC	1.52	1	1.52	0.33	0.5729	
AD	0.85	1	0.85	0.19	0.6723	
BC	1.62	1	1.62	0.35	0.5606	
BD	1.82	1	1.82	0.40	0.5381	
CD	0.31	1	0.31	0.068	0.7979	
A ²	271.60	1	271.60	59.40	< 0.0001	
B ²	682.77	1	682.77	149.33	< 0.0001	
C ²	377.64	1	377.64	82.59	< 0.0001	
D ²	708.67	1	708.67	155.00	< 0.0001	
Residual	68.58	15	4.57			
Lack of Fit	40.14	10	4.01	0.71	0.7018	not significant
Pure Error	28.44	5	5.69			
Cor Total	3748.06	29				

nanoparticles were anchored on the surface of g-C₃N₄ nanosheets after combining these three nanomaterials and formation ternary nanocomposite. The EDS spectrum of g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite was represented in Fig. 2c, which reveals the presence of Ti, O, C, N, Cu, and Fe elements without any impurities in the synthesized nanocomposite.

The magnetic properties of g-C₃N₄/TiO₂/CuFe₂O₄ were investigated by VSM analysis. The VSM data (Fig. 2d) showed that the nanocomposite exhibits superparamagnetic behavior at room temperature with a weak ferromagnetic component. The magnetization (M) values of the nanocomposite increased as the magnetic field (H) was increased, indicating that the nanocomposite has ferromagnetic characteristics. At a magnetic field of 10 kOe, the saturation magnetization (M_s) value of g-

C₃N₄/TiO₂/CuFe₂O₄ was found to be 5.62 emu/g.

UV-Vis DRS analysis was used to investigate the optical properties of the as-prepared samples and the obtained results were exhibited in Fig. 3a. Pure TiO₂ was exhibited absorption in the ultraviolet region with the absorption edge of 390 nm. However, the absorption edge of g-C₃N₄, and CuFe₂O₄ were revealed at 450 and 880, which possess intensive absorption in the visible light region. Comparing to TiO₂, the ternary g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite was exhibited a red shift, which signify the final nanocomposite can response to visible light. The band gap energies of the as-synthesized samples were calculated using Tauc plot based on Kubelka-Munk function. According to Fig. 3b, the band gap energies of g-C₃N₄, TiO₂, CuFe₂O₄, and g-C₃N₄/TiO₂/CuFe₂O₄ were obtained 2.68, 3.15, 1.28 eV, and 2.45, respectively. Therefore, as a result g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite with intensive visible light absorption can form abundant photogenerated carriers, which lead to improve the photocatalytic performance.

3.2. The model fitting and statistical analysis

To optimize the photocatalytic activity of g-C₃N₄/TiO₂/CuFe₂O₄ for degradation of TC, four main influencing variables (dosage of photocatalyst, initial concentration of TC, reaction time and solution pH) were

Table 4
Statistical parameters for the proposed quadratic model by ANOVA.

R ²	0.9799
Adj. R ²	0.9722
Pred R ²	0.9417
C. V. %	2.47
PRESS	218.60

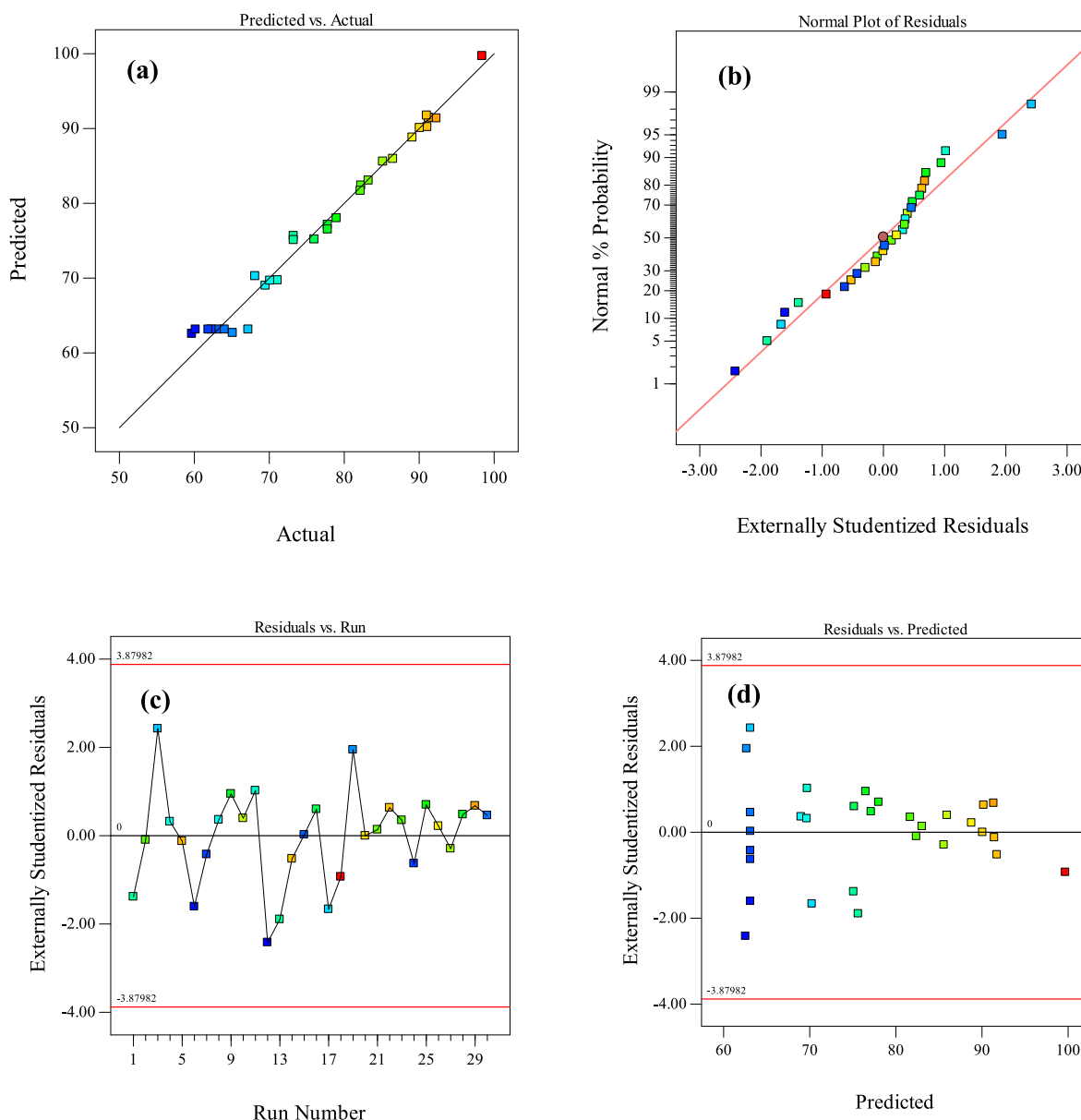


Fig. 4. Graphical plot of predicted versus actual values (a), normal probability plots of the residuals (b), externally studentized residuals versus predicted (c), and residuals versus run number (d) of the fitted model for photodegradation of TC by $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ nanocomposite.

optimized using RSM based CCD. The analysis of variance (ANOVA) was performed to investigate the adequacy and validity of the model and the results was shown in Table 3. Based on the analysis of variance (ANOVA) findings, it is evident that the proposed model and its constituent terms, namely A, B, C, D, A^2 , B^2 , C^2 , and D^2 , exhibit statistical significance, as indicated by their respective p-values being below 0.05, at a confidence level of 95%. Additionally, the p-value of the lack of fit is 0.7018, which demonstrate the lack of fit is not significant and the suggested model is adequate for predicting the TC degradation efficiency. As shown in Table 4, the high values of R^2 (0.9799), pred. R^2 (0.9722), and adj. R^2 (0.9417) show the excellent fitness between the predicted and experimental result. Further, with the reduce in the difference between R^2 , pred. R^2 , and adj. R^2 values, the entering chance non-significant terms in the suggested model is declined. Moreover, the coefficient of variation (C.V. %) value is low (2.47), which demonstrate good reliability of the experimental data. Also the value of predicted residual error sum of square (PREES = 218.60) is low reveals that the proposed model can successfully fit the data. The RSM results proposed the following second-order polynomial model in terms of coded factors:

$$\%R = 63.1333 + 6.68625 A + -4.07708B + 3.70458C + -4.12208 D + 3.14677 A^2 + 4.98927 B^2 + 3.71052 C^2 + 5.08302 D^2$$

The actual against predicted responses plot (Fig. 4a) was shown the closeness of actual values to predicted values. Therefore, the model can successfully predict the degradation efficiency of TC. The residual graphs in various modes containing normal probability plot, residuals vs. run number, and residuals vs. predicted were applied to investigate the adequacy of the model and the results was shown in Fig. 4b–d. As per the normal probability plot analysis, it is observed that the data points closely conform to a straight line, indicating that the residuals follow a normal distribution. (Dan et al., 2021). Additionally, the data points in the residuals vs. run number, and residuals vs. predicted plots are randomly scattered without any pattern. These results confirm the adequacy of the proposed model.

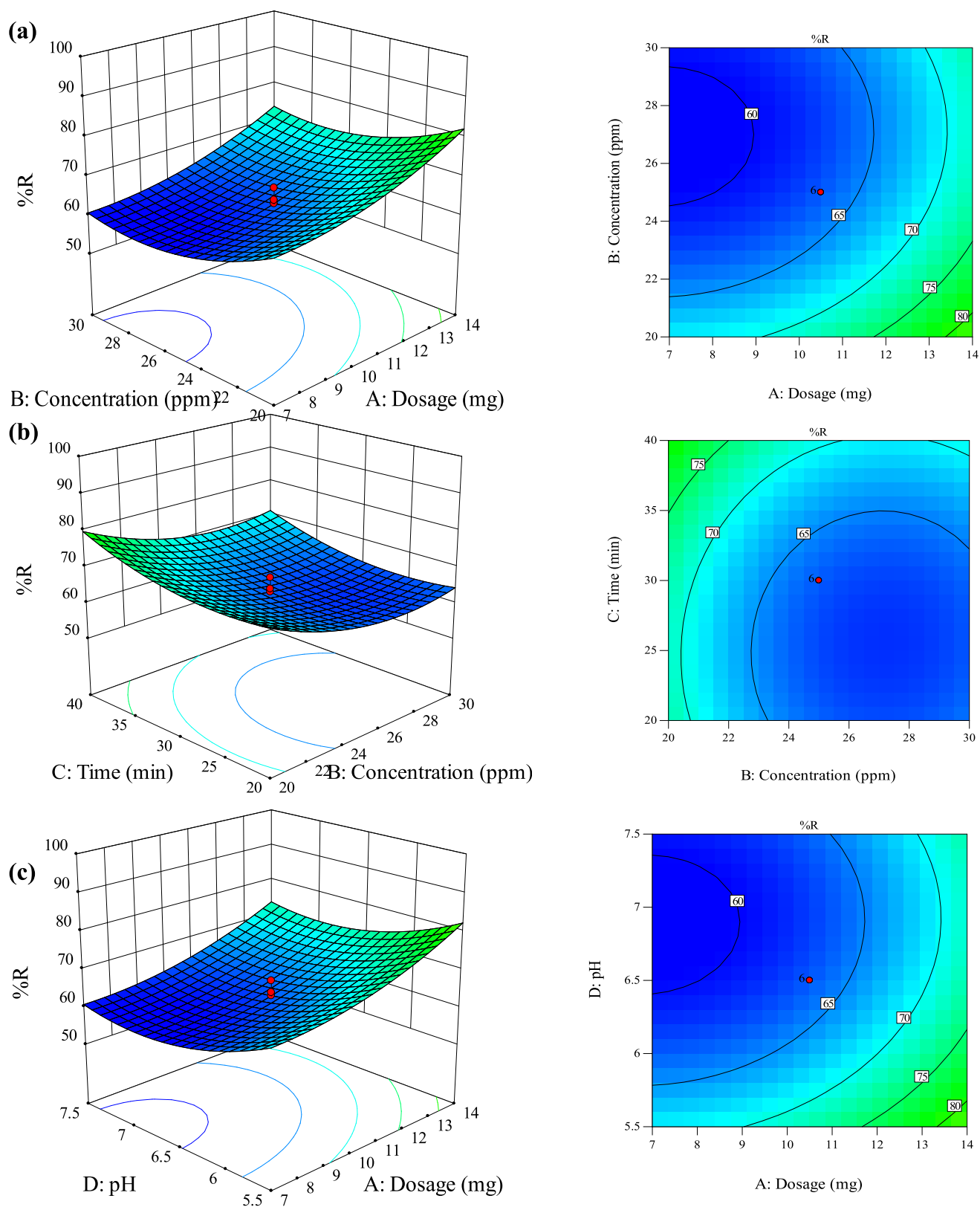


Fig. 5. Surface (3D) and contour (2D) plots for TC photodegradation by $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ nanocomposite.

3.3. Response surface analysis

The surface and contour plots were applied to visualize the effect of interactions of factors on the degradation of TC and the obtained results

were depicted in Fig. 5. The interaction between TC concentration and photocatalyst dosage was presented in Fig. 5a. The results were shown a considerable enhancement in the photocatalytic degradation efficiency of TC with the rise in the photocatalyst dosage. This enhancement trend

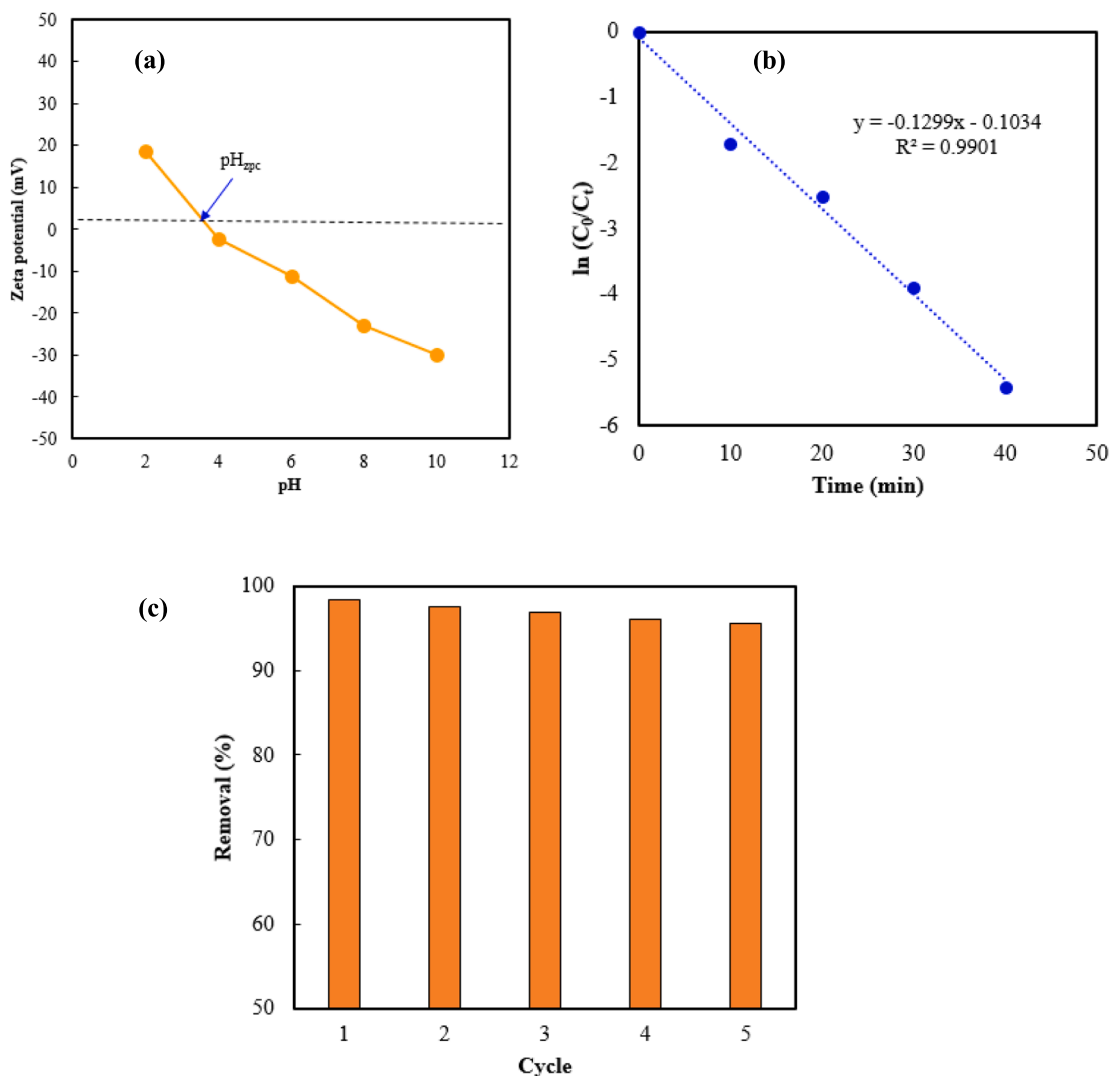


Fig. 6. Zeta potential of g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite (a), the corresponding kinetic of TC photodegradation (b), and Comparison of photocatalytic performance of various samples (c).

might be due to the increase in the active sites and adsorption capacity, which increase the photodegradation of adsorbed TC. Also, it is observed in this Figure that increasing in the initial concentration of TC lead to decline in the degradation efficiency due to unavailability of active sites and radical species in the high TC concentration. The influence of dosage of photocatalyst and time of irradiation was presented in Fig. 5b. The results demonstrate that the photocatalytic degradation percentage was increased by enhancement in the irradiation time, because increasing in the time leads towards the completion of degradation reaction. In fact, by increasing the irradiation time, the photocatalyst and pollutant have more opportunity to participate in the degradation reaction. Fig. 5c was shown the effect of TC concentration and solution pH. Regarding the effect of the pH of solution, it is clear that the degradation efficiency was increased with the increase of solution pH from 4.5 to 5.5 and at higher values, the degradation efficiency was decreased. It can be explained by zeta potential of g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite (Fig. 6a) and TC species distribution at different pHs of solution. The pH_{zpc} of g-C₃N₄/TiO₂/CuFe₂O₄ nanocomposite was 3.79, revealing that nanocomposite surface was positively charged at pH more than 3.79, and negatively charged at pH less than 3.79 [37]. On the contrary, the prevalent form of TC within the pH range of 3.3 to 7.7 predominantly consists of TC⁰ molecules. This particular pH range facilitates conditions where the interaction between the nanocomposite material and TC molecules

Table 5
Results of RSM optimization.

Response	Parameter				Predict response value	Experimental response value
	A (mg)	B (ppm)	C (min)	D		
%R	13.99	20.01	39.95	5.5	98.47	98.08

experiences reduced adsorption resistance. Consequently, this weakening of adsorption resistance enhances the availability and accessibility of active sites located on the surface of the nanocomposite. These active sites, now more readily available, effectively catalyze the degradation reactions of TC molecules. Thus, the pH-dependent predominance of TC⁰ molecules contributes significantly to the facilitation of degradation processes on the surface of the nanocomposite material. At the pH > 7.7, predominant TC species contain TCH⁻ and TC²⁻ and the degradation efficiency was declined due to charge repulsion between nanocomposite with negatively charged and molecules of TC [37].

Typically, the final step in experimental design is optimization of conditions. In this study, as shown in Table 5, the degradation efficiency of 98.47 % was suggested by the software under optimum conditions (photocatalyst dosage of 13.99 mg, TC concentration of 20.01 mg.L⁻¹ and at solution pH of 5.5 and during the course of 39.95 min reaction

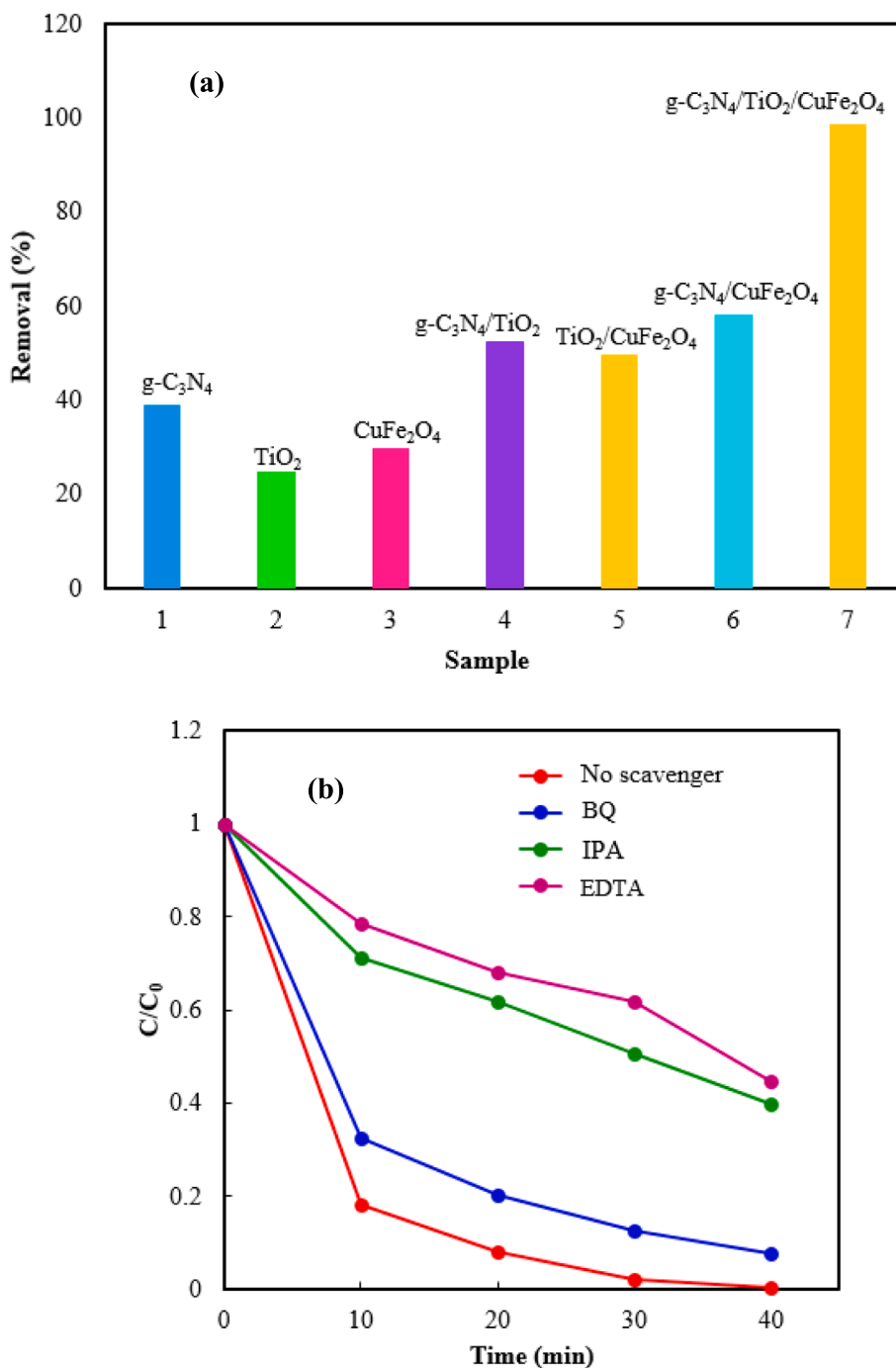


Fig. 7. Effect of various scavengers on TC photodegradation over g-C₃N₄/TiO₂/CuFe₂O₄.

time). Performing an experiment under inclusion of optimal conditions yielded the TC photodegradation efficiency of 98.08 %, being in good assert with the predicted photodegradation percentage and revealing the accuracy of the suggested RSM model.

3.4. Kinetics study

The kinetic study of the involved photocatalytic process was performed based on Langmuir-Hinshelwood's model. According to this model, the heterogeneous photodegradation process rate follows pseudo-first-kinetic model:

$$\ln(C_0/C_t) = k_{app} t$$

In which C_0 (mg.L⁻¹), and C_t (mg.L⁻¹) indicate the initial concentration of TC and the concentration of TC after light illumination, respectively. Also, k_{app} (min⁻¹) and t (min) refer to apparent rate constant and the time of light illumination. Fig. 6b was shown the kinetic results considering to g-C₃N₄/TiO₂/CuFe₂O₄ photocatalysts. As Fig. 6b presents, the degradation of TC follows the pseudo-first-kinetic model. Furthermore, the k_{app} values for g-C₃N₄/TiO₂/CuFe₂O₄ was obtained 0.1013 min⁻¹.

3.5. Stability of g-C₃N₄/TiO₂/CuFe₂O₄

In addition to desirable phtocatalytic degradation efficiency, the photocatalyst stability is important criterions to appraise a

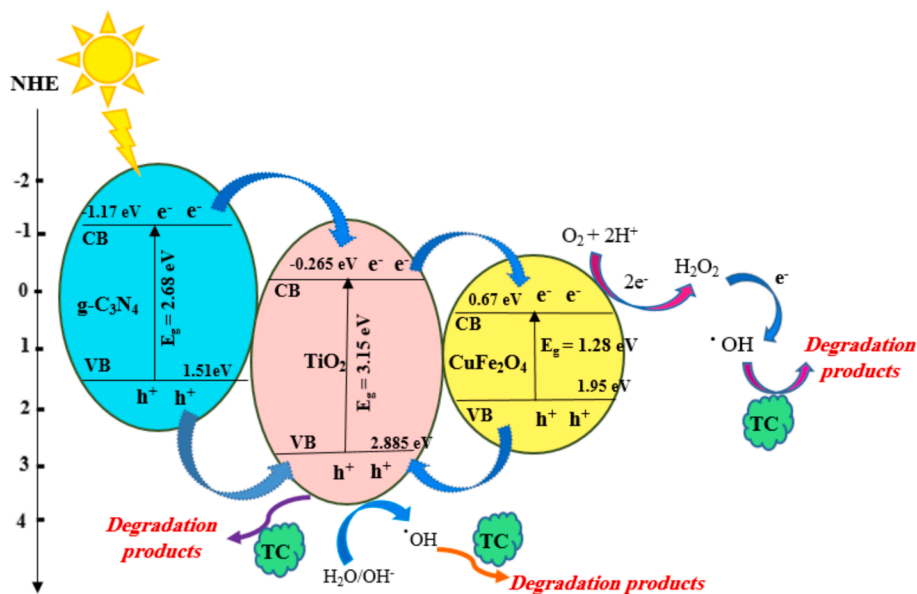


Fig. 8. Schematic diagram of TC photodegradation mechanism over $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ under visible light irradiation.

photocatalyst. Therefore, the reusability of the $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ nanocomposite was evaluated for 5 cycles under optimum conditions using circulation experiments. As shown in Fig. 6c, no significant decrease was observed in the photocatalytic efficiency after 5 cycles, which confirm the excellent stability of $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ nanocomposite.

3.6. Comparison of photocatalytic activity of various samples

The photocatalytic performance of $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ was compared with TiO_2 , $g\text{-C}_3\text{N}_4$, CuFe_2O_4 , $g\text{-C}_3\text{N}_4/\text{TiO}_2$, $\text{TiO}_2/\text{CuFe}_2\text{O}_4$, and $g\text{-C}_3\text{N}_4/\text{CuFe}_2\text{O}_4$ under optimal condition and the obtained results was shown in Fig. 7a. Notably, the $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ was exhibited the highest TC degradation activity after 40 min, which indicate the better photocatalytic activity of $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ compared to the TiO_2 , $g\text{-C}_3\text{N}_4$, CuFe_2O_4 , $g\text{-C}_3\text{N}_4/\text{TiO}_2$, and $\text{TiO}_2/\text{CuFe}_2\text{O}_4$. Therefore, the coupling of TiO_2 , $g\text{-C}_3\text{N}_4$, and CuFe_2O_4 and the refinement of nanocomposite structure resulted in substantial advancements in both electron-hole pair separation and visible light absorption. This was achieved through precise design and engineering, enabling efficient management of charge carrier mobility. Moreover, the optimized nanocomposite architecture facilitated enhanced light absorption across the visible spectrum, thereby ensuring increased utilization of solar energy for photocatalytic applications.

3.7. Photocatalytic mechanism

In order to determine the main reactive species responsible for the photodegradation of TC by $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$, radical trapping experiments were performed. Benzoquinone (BQ), isopropanol (IPA), and disodium ethylenediaminetetraacetic acid (EDTA-2Na) were used for trapping $\cdot\text{O}_2^-$, $\cdot\text{OH}$, and h^+ , respectively. As shown in Fig. 7b, after addition of EDTA-2Na and IPA to the solution, degradation efficiency was decreased, while addition of BQ has no significant influence on the TC photodegradation efficiency. Briefly, the radical trapping experiments reveal that the main reactive species responsible for degradation process of TC was in the order of $\cdot\text{OH} > h^+ > \cdot\text{O}_2^-$.

According to the above trapping experiments results, E_g , relative potentials of the conduction band (E_{CB}) and valance band (E_{VB}) of samples, a possible photocatalytic mechanism for photodegradation of TC by $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ was proposed and presented in Fig. 8. After

visible light absorption by photocatalyst, e^- and h^+ were generated in the CB and VB of TiO_2 , $g\text{-C}_3\text{N}_4$, and CuFe_2O_4 . According to the energy diagram (Fig. 8), the photogenerated electrons from the CB of $g\text{-C}_3\text{N}_4$ spontaneously move to the CB of TiO_2 , and then from the CB of TiO_2 move to the CB of CuFe_2O_4 . Furthermore, the holes remain in the VB of TiO_2 and can participate directly in the degradation reaction of TC. Also, the E_{VB} of TiO_2 is more positive than the standard redox potentials of $\text{OH}^-/\cdot\text{OH}$ (1.99 eV vs. NHE) and $\text{H}_2\text{O}/\cdot\text{OH}$ (2.37 eV vs. NHE), which indicate the holes can react with OH^- or H_2O to generate $\cdot\text{OH}$. The E_{CB} of CuFe_2O_4 is more positive than the standard redox potential of $\text{O}_2/\cdot\text{O}_2^-$ (-0.33 eV vs. NHE), which indicated the electrons in the CB of CuFe_2O_4 cannot reduce the O_2 on the surface of the photocatalyst to $\cdot\text{O}_2^-$ species. The CB of CuFe_2O_4 is more negative than the standard redox potential ($\text{O}_2/\text{H}_2\text{O}_2$) (0.685 eV vs. NHE). So, the electrons in the CB of CuFe_2O_4 could be trapped by absorbed O_2 and H^+ in the solution to form H_2O_2 species. Then, the H_2O_2 react with electrons and produce $\cdot\text{OH}$ species to participate in the TC photodegradation reaction.

4. Conclusion

In this study, the novel magnetic ternary $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ nanocomposite was fabricated via wet impregnation method. The as-synthesized nanocomposite was shown significantly enhanced photocatalytic activity for degradation of TC under visible light illumination. More importantly, the TC photodegradation by $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ nanocomposite has attained higher photodegradation efficiency, which is much higher than that of pristine TiO_2 , $g\text{-C}_3\text{N}_4$, CuFe_2O_4 , and binary $g\text{-C}_3\text{N}_4/\text{TiO}_2$, $\text{TiO}_2/\text{CuFe}_2\text{O}_4$, and $g\text{-C}_3\text{N}_4/\text{CuFe}_2\text{O}_4$ nanocomposites. Notably, the extended optical absorption, and high electron/hole separation improve the photocatalytic degradation efficiency of TC by $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ nanocomposite. Moreover, CCD-RSM was applied to investigate the effect of operational parameters on the TC photodegradation efficiency. Complete photodegradation of TC occurred under the conditions of 13.99 mg nanocomposite, 20.01 $\text{mg}\cdot\text{L}^{-1}$ TC, and pH 5.5, within 40 min. According to the scavenging experiments, holes and hydroxyl radicals were dominant species in the degradation process. This study indicates that the nanocomposite of $g\text{-C}_3\text{N}_4/\text{TiO}_2/\text{CuFe}_2\text{O}_4$ may be expected to be a promising candidate for removal of TC from aqueous solution.

CRediT authorship contribution statement

Khalid Mujasam Batoó: Writing – review & editing, Resources, Project administration, Investigation, Funding acquisition, Data curation, Conceptualization. **Kadhim Hussein Jassim:** Writing – review & editing, Visualization, Validation, Software, Methodology, Formal analysis, Data curation. **Talal Aziz Qassem:** Writing – review & editing, Validation, Supervision, Project administration, Funding acquisition, Data curation, Conceptualization. **Sajjad Hussain:** Writing – original draft, Software, Resources, Methodology, Investigation, Formal analysis. **Wafaa Talib Hasson:** Writing – original draft, Software, Resources, Methodology, Investigation, Formal analysis, Data curation. **Sarah Salah Jalal:** Visualization, Validation, Software, Resources, Methodology, Investigation. **Montather F. Ramadan:** Writing – original draft, Software, Resources, Methodology, Investigation, Formal analysis, Data curation, Conceptualization. **Safaa Mustafa Hameed:** Writing – original draft, Visualization, Validation, Software, Data curation, Conceptualization. **Ahmed Hussien Alawadi:** Writing – review & editing, Visualization, Validation, Software, Resources, Data curation, Conceptualization. **Ali Alsaalamy:** Writing – review & editing, Visualization, Validation, Software, Methodology, Investigation, Formal analysis, Data curation.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Appendix A. Supplementary data

Supplementary data to this article can be found online at <https://doi.org/10.1016/j.jscs.2024.101871>.

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